HYDRODYNAMICS AND BIO-OPTICAL ASSESSMENT OF TWO PRISTINE SUBTROPICAL ESTUARIES IN SOUTHERN BRAZIL

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ABSTRACT

Measuring small-scale physical processes and how they affect the spatial patterns of sea water's optical constituents plays a key role in understanding the functioning of complex coastal ecosystems such as estuaries. The spatial variations of physical and bio-optical water properties were investigated during one spring tidal cycle in the austral summer, on two transects across the channel of the Medeiros and Itaqui sub-estuaries. These sub-estuaries are a biosphere reserve, and part of the Paranaguá Estuarine System, located on the Southern Brazilian coast. Both sub-estuaries were classified as Type 1a, well-mixed with low stratification. The salinity variations are in phase with the water level, and the tidal propagation is well represented by a standing wave. The vertical velocity profiles showed little vertical shear, and the intensity of the u-component of the velocity varied semi-diurnally. The upper estuary salt transport was dominated by tidal diffusion in an unstable water column. The optical environment presented a mixed dominance of optically active substances, as indicated by the absorption coefficients of dissolved and particulate matter. The colored dissolved organic matter (CDOM) showed overall conservative behavior and was dominant in light absorption below 550 nm in the Medeiros, while non-algal particles play the most important role in light absorption in the Itaqui in the blue absorption band. The phytoplanktonic contribution is prominent in the red domain and increases as a function of saline intrusion. However, due to the influence of freshwater discharge and the re-suspension of bottom sediments induced by physical processes, the concentrations of the optical components in the water column do not generally have any simple relationships between them.

RESUMO

A quantificação de processos de transporte de pequena escala e a compreensão de como eles afetam o padrão espacial dos constituintes ópticos da água são peça chave no funcionamento de um ecossistema costeiro. A variação espacial e local de propriedades físicas e bio-ópticas da água foi analisada durante um ciclo de maré de sizígia do verão austral, bem como as estimativas do transporte advectivo e dispersivo de sal na secção transversal da boca dos subestuários Medeiros e Itaqui. Estes subestuários fazem parte do Sistema Estuarino de Paranaguá, sul do Brasil, e também integram a reserva da biosfera. Ambos os subestuários foram classificados como Tipo 1a, bem misturados com pouca estratificação. As variações da salinidade estão em fase com o nível da água, e a propagação da onda de maré se dá na forma de uma onda estacionária. Os perfis verticais de velocidade mostraram pouco cisalhamento vertical, e a intensidade da componente u de velocidade variou semidurmanente. O transporte de sal estuarino acima foi dominado pela difusão da maré em uma coluna de água instável. O ambiente óptico apresentou uma dominância mista dos componentes opticamente ativos, como indicado pelos coeficientes de absorção dos materiais dissolvidos e particulados. A matéria orgânica dissolvida colorida (CDOM) apresentou um comportamento conservativo com dominância da luz abaixo de 500 nm em Medeiros, enquanto que no subestuário Itaqui as partículas não algas foram dominantes na faixa do azul. A contribuição do fitoplâncton é alta na faixa de vermelho e aumenta em função da intrusão salina. Contudo, devido à influência da descarga de água doce e a re-suspensão de sedimentos induzida por processos físicos as concentrações dos componentes ópticos na coluna d’água geralmente não apresentam uma relação simples entre si.


Descritores: Circulação estuarina, Transporte de sal, Qualidade da água, Classificação óptica, Coeficiente de absorção, Baía das Laranjeiras.
INTRODUCTION

The Paranaguá Estuarine Complex (PEC), located on the Paraná State coast in south-eastern Brazil (lat: 25°30’S; long: 48°25’W), is a large, interconnected subtropical estuarine system that comprises two main water bodies and a wide variety of intertidal environments, such as mangrove swamps, marshes, and non-vegetated tidal flats (LANA et al., 2000). Fishing and commercial activities together with urban and tourism pressure throughout the PEC make it one of the most economically and ecologically important estuaries on the Brazilian coast.

The two main water bodies of the PEC have similar environmental characteristics but are subject to distinct anthropogenic pressures. The east-west axis is formed essentially by the Paranaguá and Antonina Bays (Fig. 1) and has been influenced by the anthropogenic input of domestic discharges, port and industrial effluents, and intense contamination with sewage near the city of Paranaguá (MARTINS et al., 2010). The north-south axis is formed by Laranjeiras Bay, an Environmental Protection Area of ecological importance. The Laranjeiras Bay catchment area is one of the last preserved Atlantic rainforests of South America and has been protected as a biosphere reserve by the United Nations Educational, Scientific and Cultural Organization (UNESCO) since 1993. The estuary can be classified as pristine, with neither anthropogenic impacts caused by the introduction of oil and its derivates nor any fecal contribution (MARTINS et al., 2012).

Fig. 1. The Paranaguá Estuarine Complex, composed of Laranjeiras Bay and Paranaguá Bay, and the delimitation of the catchment areas. Paranaguá Harbour, Paranaguá City, Paranaguá Bay, Laranjeiras Bay and Guaraqueçaba Bay are indicated by arrows (upper left). The studied areas – the Medeiros (MD) and Itaqui (IT) inlets – and the sample station (circle), anchor stations (triangle), moored station (square) and transect (line) are shown in detail at right. The morphological features of the cross sections are show at bottom left.
The PEC region has a subtropical wet climate, with distinct wet and dry seasons. The mean annual rainfall is 2500 mm (maximum 5300 mm), and the mean annual relative humidity is approximately 85%. The mean precipitation during the rainy season (summer: December to March) is more than three times higher than during the dry season (winter). The tidal regime is semidiurnal with diurnal inequalities. Mean neap and spring tidal heights are, respectively, 1.3 and 1.7 m at the bay mouth and 2.0 and 2.7 m at Antonina (MARONE and JAMIYANAA, 1997). Stratification and mixing processes in the main channels are controlled primarily by tidal currents and secondarily by freshwater discharge, which causes seasonal variations in the magnitude of vertical salinity stratification. Paranaguá Bay can exhibit highly stratified (summer neap tides), partly mixed (summer spring tides) and well-mixed (winter spring tides) conditions within a single season or fortnight (MANTOVANELLI et al., 2004). Despite its great environmental importance and preservation status, Laranjeiras Bay is under-studied; information on the hydrographic and hydrodynamic characteristics of its sub-estuaries is practically non-existent, or at least unpublished. NOERNBERG et al. (2006) documented the morphological and hydrological characteristics of this region. The main characteristics of the two sub-estuaries that are the focus of this study are presented in Table 1.

Table 1. Morphological and hydrological characteristics of the Medeiros and Itaqui sub-estuaries (adapted from NOERNBERG et al., 2006).

<table>
<thead>
<tr>
<th></th>
<th>Medeiros</th>
<th>Itaqui</th>
</tr>
</thead>
<tbody>
<tr>
<td>Drainage area (km²)</td>
<td>44.6</td>
<td>122.8</td>
</tr>
<tr>
<td>Drainage density (river/km²)</td>
<td>0.98</td>
<td>1.88</td>
</tr>
<tr>
<td>Hydraulic gradient</td>
<td>0.07</td>
<td>0.04</td>
</tr>
<tr>
<td>Maximum width to length ratio</td>
<td>0.14</td>
<td>0.27</td>
</tr>
<tr>
<td>Wetlands area (km²)</td>
<td>9.1</td>
<td>8.6</td>
</tr>
<tr>
<td>Water body area (km²)</td>
<td>4.12</td>
<td>7.74</td>
</tr>
<tr>
<td>Tidal flat area (km²)</td>
<td>2.6</td>
<td>4.2</td>
</tr>
<tr>
<td>Relative tidal flat area (%)</td>
<td>63.1</td>
<td>54.3</td>
</tr>
</tbody>
</table>

There is increasing interest in describing and assessing the environmental conditions along the north-south axis of the PEC. Environmental damage related to sporadic harmful algae blooms (FERNANDES et al., 2001; MAFRA JR et al., 2006), oil spills and the discharge of polluted effluents by a shrimp farm and small villages has already been observed in the area. These negative impacts are neither quantified nor qualified due to the lack of an environmental monitoring program and the ineffectiveness of management plans for the coastal region of the state of Paraná. Therefore, assessing the environmental setting, such as the hydrodynamics and bio-optical properties, of these pristine estuaries is of the utmost importance to improve the understanding of these complex coastal environments and to develop adequate management strategies for the region.

Quantifying and understanding the natural processes that occur, on a relatively short time scale, in coastal environments such as estuaries, is not only relevant within a scientific perspective, but also crucial for the planning and management of these valuable natural resources. Reliable data are invaluable for decision makers’ identification of the reasons for occurring environmental problems (JL, 2008).

The aim of this study is to investigate the spatial and local variations of physical and bio-optical water properties of the two sub-estuaries of the PEC (the Medeiros and the Itaqui). Specific objectives of this study are: i) the estimation of the advective and dispersive salt transport along fixed transects across the channels of the Medeiros and Itaqui sub-estuaries during a spring tidal cycle; ii) the quantification of the relative contributions of the absorption of phytoplankton (aₚₚ), non-algal particles (aₐₐₐ), and colored dissolved organic matter (aₐₐₐₐ) to the total absorption coefficient. The relationship of these variables to salinity is assessed to investigate the possible use of color images as proxies of water circulation through variations in the salt content.

Our approach was to describe the basic thermohaline, hydrodynamic and bio-optical characteristics of the two sub-estuaries of the PEC, aiming to set the stage for the long-term goal of establishing a physical-biological-geochemical-optical platform combining modelling, remote sensing and in situ observations to guide the management of this coastal ecosystem.

**Material and Methods**

Surveys were conducted in March, 2012 on the Medeiros sub-estuary (day 05) and the Itaqui sub-estuary (day 09). Tidal height, current speed and direction, and hydrographic properties were measured by an InterOcean S4-CTD electromagnetic current meter which collected data every fifteen minutes, with a frequency of acquisition of 2 Hz, over one spring tidal cycle at a moored station positioned one meter above bottom at the sub-estuary mouth (Fig. 1). Current velocity profiles were measured hourly along a fixed transect across the mouth of each estuary during a 13–hour tidal cycle, using a boat-mounted 1000 kHz Sontek ADP. The instrument was set up with 0.8 m bin resolution and 1 minute averaging.
Between every ADP transect the vertical profiles of hydrographic properties were measured with an anchored Alec CTD, model ASTD. The vertical profiles of hydrographic properties were also measured at hourly intervals at four and three stations, respectively, along each of the sub-estuaries (of the Medeiros and the Itaqui). Water samples from the subsurface (approximately 1 m depth) were taken at least once during each phase of the tidal cycle (high, ebb, low and flood tide) to determine the concentrations of chlorophyll and particulate suspended matter and the absorption coefficients of dissolved organic and particulate matter.

The hydrographic properties such as the temperature, conductivity and pressure (T, C and p) were converted into salinity (S) and sigma-t (σ_t), in accordance with the Practical Salinity Scale (PSS-78) and the International Equation of State of Sea Water (IES-80). The velocity vector was decomposed into axial (u-component) and transverse (v-component) components in accordance with Oxyz, the local referential system, with the origin on the free surface of the water. Ox and Oy were the longitudinal and transversal axes, oriented positively seawards and to the left margin respectively, and the vertical axis, Oz, was oriented against gravity acceleration (positive values upward).

Due to the shallowness of both channels’ cross-sections (depths between 4.2 and 5.7 m for the Medeiros and 4.4 and 6.6 m for the Itaqui at the slack low and high water respectively), the sampling depth (z) was normalised according to KJERFVE (1975) by its dimensionless value \( Z = z / h(t) \), where \( h(t) \) is the water depth at the sampling time \( t \). All the measurements performed at each sub-estuary cross-section were quality controlled, smoothed and interpolated along the water column at intervals of 0.05 \( Z \).

The sub-estuaries were classified in accordance with the classical stratification-circulation diagram theoretically derived by Hansen and Rattray (1966), which was also used to calculate the relative contributions of the advective and diffusive processes to the upstream salt transport through its key parameter \( v \). The advective salt transport components, such as the river discharge, tidal diffusion, Stokes drift, and density currents, were calculated according to a methodology adapted from Miranda et al. (2002) on the assumption that the channel was laterally homogeneous.

The stability characteristics of the sub-estuaries during the spring tidal cycle were analysed using the Richardson layer number \( R_{iL} \) (Bowden, 1963, 1978), where \( h(t) \) and \( \Delta \rho_s(t) \) are the local depth (m), the bottom density minus the surface density, the mean-depth density (kg m\(^{-3}\)) and the mean-depth velocity (m s\(^{-1}\)), respectively.

The physical interpretation of the layer Richardson number depends on its magnitude: for \( R_{iL} > 20 \), the water column is highly stable with low vertical mixing; for \( R_{iL} < 20 \), the bottom turbulence causes vertical mixing in the water column; and below the critical value \( R_{iL} \approx 2 \), turbulent mixing makes the water column unstable (Dyer and New, 1986).

Water samples were collected during four tidal phases in each sub-estuary (Table 2) using dark plastic bottles that were pre-washed with 10% HCl, purified water and rinsed with local water before being stored in a refrigerated container. The water samples were stored in dark plastic bottles inside a cooling box until processing that occurred after the sampling campaigns. Due to the lack of logistic infrastructure on the survey platforms, the samples were kept as stable as possible for a maximum of 12 hours before filtration. The samples were filtered in accordance with the methodology described by the NASA protocols (Mitchell et al., 2002) for the determination of the absorption coefficients of colored dissolved organic matter (CDOM) and particulate matter. Whatman GF/F filters were used to determine the suspended particulate matter and chlorophyll concentrations following the gravimetric and fluorometric methodologies, respectively (Strickland and Parsons, 1972). Chlorophyll concentrations were determined for the 24 h acetone (90%) extraction sample using a Turner Design Fluorometer (10AU).

<table>
<thead>
<tr>
<th>Time</th>
<th>Tide</th>
<th>Reference</th>
<th>Time</th>
<th>Tide</th>
<th>Reference</th>
</tr>
</thead>
<tbody>
<tr>
<td>08:23</td>
<td>Low</td>
<td>M1</td>
<td>7:15</td>
<td>Ebb</td>
<td>I1</td>
</tr>
<tr>
<td>11:05</td>
<td>Spring</td>
<td>M2</td>
<td>9:30</td>
<td>Low</td>
<td>I2</td>
</tr>
<tr>
<td>15:25</td>
<td>High</td>
<td>M3</td>
<td>12:05</td>
<td>Spring</td>
<td>I3</td>
</tr>
<tr>
<td>20:10</td>
<td>Ebb</td>
<td>M4</td>
<td>15:27</td>
<td>High</td>
<td>I4</td>
</tr>
</tbody>
</table>

The light absorption coefficient for CDOM was measured in 0.2 µm mixed membrane filters that were prewashed for 15 minutes with 5% HCl before filtering. The seawater filtering was performed in four steps: initially three successive filtrations of 50 ml to wash the filter and filtering apparatus, followed by a fourth filtration of approximately 400 ml, from which 50 ml of the CDOM sample were preserved in an amber glass bottle (prewashed with 10% HCl and rinsed 3 times) and 350 ml used in the particulate matter absorption coefficient analysis. After filtration,
the CDOM absorbance (250-800 nm, 1 nm step) was measured with a spectrophotometer (Shimadzu UV-1601PC) using ultra-filtered distilled water as a blank. The absorbance was transformed into absorption coefficients (dimensions m$^{-1}$) by subtracting the blank, normalising it at 660 nm (BRICAUD et al., 1981), multiplying it by 2.303 and dividing it by the path length through the spectrophotometer cuvette (0.1 m).

The absorption coefficients of particulate matter were determined using GF/F glass-fiber filters (Whatman). Different volumes were filtered in dim light, and the filtering apparatus was rinsed with the filtered seawater to gather the retained material into the filter and to eliminate the salt content. The method followed the “quantitative filter technique” described by MITCHELL et al. (2002). First, the absorbance of the total particulate matter (300-750 nm, 1 nm step) was determined with the spectrophotometer, and the filter was then soaked in hot methanol for 90 minutes to extract all the pigments. This ‘bleached’ filter was rinsed with the filtered seawater to eliminate the methanol. Finally, the absorbance of the remaining material, termed non-algal particles (NAP or detritus), was measured. The absorbance of the particulate matter and of the NAP were converted into absorption coefficients by using the averaged values between 740-750 nm as a baseline, subtracting the absorbance of the blank filter, multiplying them by 2.303, and dividing them by the $\beta$-scaling factor (calculated according to MITCHELL et al., 2002) and the geometric absorption path length (the volume of filtered liquid multiplied by the area of the filter). The difference between the absorption coefficient of the total matter and NAP gave the absorption coefficient of phytoplankton.

**RESULTS**

January and March are climatologically the rainiest months (DNM, 1992), but they exhibited lower-than-average rainfall in 2012. In January and March, 2012 the rainfall at Paranaguá was 34.2% and 86.7%, respectively, lower than the twenty-nine year average (Fig. 2).

**Currents and Hydrography**

The bottom salinity variations correlate with tidal height at the bottom moored station, with the maximum salinity occurring close to high water on both the Medeiros and Itaqui sub-estuaries (Figs 3a and 3b). On the Itaqui, an oscillation in salinity occurs close to low water, which is likely associated with the complex transversal bathymetry (Fig. 1).
The hourly vertical profiles of salinity (Figs. 5a and 5b) show the salinity variations caused by the advective influence of tidal currents and the background river discharge. On the Medeiros, the average salinity was 23.56 ups at the surface and 24.63 ups at the bottom. On the Itaqui, the average salinity was 21.56 ups at the surface and 23.75 ups at the bottom. Over the tidal cycle water temperatures varied from 28.4°C to 30.3°C on the Medeiros and 28.6°C to 30.2°C on the Itaqui, with the highest temperatures occurring during the time of greatest solar radiation. The vertical distribution of temperature presented nearly isothermal conditions with mean variations of 0.7°C and 0.5°C for the Medeiros and the Itaqui, respectively.

The vertical velocity profiles showed little vertical shear (Fig. 4) in either of the sub-estuaries, except near the bottom as a result of friction. On the Medeiros, the buoyancy was less important in controlling the velocity shear because of the weaker vertical salinity gradients (Fig. 5a). On the Itaqui, the salinity profiles indicate a shallow halocline erosion and intensification (Fig. 5b); the maximum surface and bottom salinities occurring close to high water, when the water column presents some degree of stratification due to river discharge.

The water column stability was evaluated using the Richardson layer number ($R_i$). Both sub-estuaries showed dominance of turbulent flows ($R_i < 2$) during the full tidal cycle, however, on the Itaqui it was possible to observe short periods (~ 2 h) of weak stability ($2 < R_i < 20$) during both low and high tides.

**Table 3. Main advective salinity transport components per unit width for the spring tidal cycle in the Medeiros and Itaqui sub-estuaries.**

<table>
<thead>
<tr>
<th>Components</th>
<th>Medeiros</th>
<th>Itaqui</th>
</tr>
</thead>
<tbody>
<tr>
<td>Residual velocity (river input)</td>
<td>1.72</td>
<td>1.84</td>
</tr>
<tr>
<td>Stokes drift</td>
<td>0.20</td>
<td>0.44</td>
</tr>
<tr>
<td>Tidal diffusion</td>
<td>0.04</td>
<td>0.04</td>
</tr>
<tr>
<td>Baroclinic forcing</td>
<td>0.00</td>
<td>0.00</td>
</tr>
<tr>
<td>S</td>
<td>1.96</td>
<td>2.32</td>
</tr>
</tbody>
</table>

units in kg m$^{-1}$ s$^{-1}$

The stratification ($\rho_e$) and circulation parameters ($\rho_c$) were calculated using the time-averaged profiles of salinity and of the $u$-component of velocity. A description of these parameters can be
found in HANSEN and RATTRAY (1966). These parameters are shown in the composite stratification-circulation diagram for the Medeiros and Itaqui sub-estuaries (Fig. 6). Both sub-estuaries were classified as Type 1a (well-mixed and low stratification), with ν=0.99. The sub-estuaries were characterised by a unidirectional time-averaged velocity profile.

Bio-Optical Parameters: Chlorophyll and Suspended Particulate Matter Concentrations and Dissolved and Particulate Absorption Coefficients

Chlorophyll-$a$ concentrations ranged from 6 to 12 mg·m$^{-3}$, while suspended particulate matter (SPM) varied over one order of magnitude. A maximum of 22.3 g·m$^{-3}$ of SPM was recorded at the first low tide sampling from the Medeiros (sample M1), along with a maximum of 12.6 mg·m$^{-3}$ of chlorophyll-$a$, whereas SPM remained under 2 g·m$^{-3}$ during the rest of the tidal cycle at this location. Maxima chlorophyll (11.1 mg·m$^{-3}$) and SPM (13.2 g·m$^{-3}$) values were also concomitantly measured during ebb tide in the Itaqui sub-estuary (sample I1), however the concentrations here had less variation than that observed in the Medeiros. Average concentrations of the bio-optical parameters, absorption coefficients and exponential slopes are presented in Table 4.

The CDOM spectral absorption coefficients ($a_{CDOM} (\lambda)$) followed exponential functions, with higher $a_{CDOM} (\lambda)$ observed in the Medeiros sub-estuary and during low tide at both sub-estuaries. Figure 7 presents $a_{CDOM} (\lambda)$ and the average exponential slope ($S_{CDOM}$) calculated following BABIN et al. (2003) for absorption coefficients at 350-500 nm. The dominance of the conservative behaviour of CDOM is supported by the strong correlation between $a_{CDOM} (\lambda)$, at 250-450 nm, and salinity at the two locations (Fig. 8). The spectral slope ratio $S_R$ (defined as $S_{(275-295)} / S_{(350-400)}$), used here to indicate the source of the dissolved organic matter (DOM), according to the relation between $a_{CDOM} (\lambda)$ and the molecular weight of DOM (HELMS et al., 2008), showed little variation from its average of 0.95 (standard deviation 0.03) for both sub-estuaries. The high correlation between $a_{CDOM} (\lambda)$ and salinity together with $S_R < 1$ provide strong indications that terrestrial inputs dominate the dissolved organic pool in the Medeiros and Itaqui sub-estuaries.
Table 4. Average values and standard deviations of bio-optical parameters recorded in the Medeiros and Itaqui sub-estuaries. Bracketed values give the observed ranges (minima and maxima).

<table>
<thead>
<tr>
<th>Variable</th>
<th>Medeiros</th>
<th>Itaqui</th>
</tr>
</thead>
<tbody>
<tr>
<td>[Chla] (mg·m⁻³)</td>
<td>8.5±2.8 [6.7-12.6]</td>
<td>9.1±1.9 [7.0-11.1]</td>
</tr>
<tr>
<td>[SPM] (g·m⁻³)</td>
<td>1.4±0.4* [1.1-1.8]</td>
<td>8.0±5.0 [1.9-13.2]</td>
</tr>
<tr>
<td>α_CDOM (443)</td>
<td>0.498±0.146 [0.355-0.652]</td>
<td>0.352±0.077 [0.279-0.456]</td>
</tr>
<tr>
<td>S_CDOM (350-550)</td>
<td>0.0171±0.0009 [0.0164-0.0183]</td>
<td>0.0181±0.001 [0.0166-0.0187]</td>
</tr>
<tr>
<td>S_R (275-295;350-400)</td>
<td>0.94±0.03 [0.91-0.97]</td>
<td>0.97±0.05 [0.92-1.03]</td>
</tr>
<tr>
<td>α_NAP (443)</td>
<td>0.316±0.056 [0.253-0.387]</td>
<td>0.569±0.332 [0.269-0.884]</td>
</tr>
<tr>
<td>S_NAP (380-730)**</td>
<td>0.0130±0.0005 [0.0126-0.0137]</td>
<td>0.0128±0.0006 [0.0123-0.0137]</td>
</tr>
<tr>
<td>α_Ph (443)</td>
<td>0.233±0.016 [0.213-0.249]</td>
<td>0.173±0.043 [0.122-0.216]</td>
</tr>
<tr>
<td>a°_Ph (443)</td>
<td>0.030±0.008 [0.017-0.036]</td>
<td>0.019±0.001 [0.017-0.020]</td>
</tr>
<tr>
<td>α_NAP (675)</td>
<td>0.099±0.006 [0.093-0.106]</td>
<td>0.101±0.022 [0.077-0.124]</td>
</tr>
</tbody>
</table>

*Outlier of 22.4 g·m⁻³ excluded. ** Exponential slope calculated following Babin et al. (2003) using the absorption coefficients between 380-730 nm but excluding 400-480 nm and 620-710 nm.

Fig. 7. Time variation of hourly vertical profiles of the u-component of velocity at the anchor station at Medeiros (a) and Itaqui (b). u<0 and u>0 indicates up and down estuary current directions.
The particulate matter absorption coefficient \( a_p(\lambda) \) combines the effects of phytoplankton \( (a_{ph}) \) and remaining non-algal particles \( (a_{NAP}) \) on light attenuation within the water column. NAP is a mixture of inorganic material and living or non-living heterotrophic organisms, while the contribution of phytoplankton can be distinguished mainly by its photosynthetic pigment signatures (especially chlorophyll-\(a\)). The absorption spectra of the particulate matter in the Medeiros and Itaqui sub-estuaries (Fig. 9) showed similar features overall, with the exception of the ebb- and low-tide phases in the Itaqui sub-estuary. These two tidal phases were characterised by high values of \( a_p(\lambda) \), mainly due to an increase in the NAP contribution. The NAP absorption coefficients showed an exponential signature along the visible domain, with \( S_{NAP} \) varying between a maximum of 0.0137 nm\(^{-1}\) during flooding tide in the Medeiros and high tide in the Itaqui and a minimum of 0.0123 nm\(^{-1}\) at low tide in the Itaqui. These values were calculated from the slope of the exponential fit from 380 to 730 nm, excluding the 400-480 nm and 620-710 nm bands, following Babin et al., 2003. Absorption by NAP was predominantly by particles at 443 nm in both sub-estuaries (Fig. 10). In the Itaqui the contribution of detritus to light attenuation reached maxima of 80% and 87% of \( a_p(\lambda) \) during ebb and low tides, respectively, whereas maxima relative contribution of phytoplankton to \( a_p(\lambda) \) accounted for only 50%.
The relative contributions of $\alpha_{\text{NAP}}(\lambda)$ (light grey columns) and $\alpha_{\text{Ph}}(\lambda)$ (dark grey columns) to $\alpha_{\text{p}}(\lambda)$ (black columns) at 443 nm, and the ratio (%) of $\alpha_{\text{Ph}}(\lambda)$ to $\alpha_{\text{p}}(\lambda)$ during tidal phases at the Medeiros and Itaqui inlets.

The spectral signatures of chlorophyll-\textit{a} near 443 nm and 675 nm are clearly to be observed along the absorption profiles of phytoplankton from both sub-estuaries (Fig. 11). A shoulder can be seen at approximately 625-635 nm (Fig. 11). Strong positive correlations between $\alpha_{\text{Ph}}(443)$ and chlorophyll concentration ($r^2$=0.99) and between $\alpha_{\text{NAP}}(443)$ and the concentration of SPM ($r^2$=0.94) were found in the Itaqui, while an inverse correlation between $\alpha_{\text{Ph}}(443)$ and chlorophyll concentration ($r^2$=-0.86) was found in the Medeiros. Even though this inverse relationship is highlighted, it is important to note that the small dispersion of four chlorophyll and absorption records may not be enough to characterize a consistent situation. The maxima SPM concentration recorded in Sample M1 masked the positive relationship with $\alpha_{\text{NAP}}(443)$ that is evidenced when the SPM outliers are excluded from the fit ($r^2$=0.99). Still, the elevated NAP contribution to absorption in water is related to lower salinities; the tidal intrusion can also be related to a higher phytoplanktonic contribution to $\alpha_{\text{p}}(443)$. The specific phytoplankton absorption coefficient ($\alpha_{\text{Ph}}^*$ at 443 nm) ranged from a minimum of 0.017 m$^2$·mg$^{-1}$ for the Itaqui to a maximum of 0.036 in the Medeiros sub-estuary.

The ternary diagrams (Fig. 12) indicate a shift from CDOM and NAP dominance at 443 nm and 555 nm to phytoplankton at 675 nm. It is possible to verify that CDOM dominated the absorption in the Medeiros sub-estuary at smaller wavelengths, whereas the NAP contribution is more evident at Itaqui.
Observations of physical and bio-optical properties of the estuarine water column, obtained during the spring tidal cycles for the Medeiros and Itaqui sub-estuaries at the PEC, were planned for March, a season of climatologically wet conditions. The surveys were performed in March 2012, a year when the austral summer, especially March, was around 30% drier than normal. Both sub-estuaries were well-mixed with low stratification (Type 1a) during the surveys, with $v=0.99$. The larger river, the Itaqui, has a drainage area three times greater than and a drainage density twice that of the Medeiros, and in normal (wet) summer conditions, is likely to present a partially mixed, Type 2, classification, as observed at Paranaguá Bay by MANTOVANELLI et al. (2004) and at Bertioga (SP) by MIRANDA et al., (1998).

Salinity variations measured at a moored station were in phase with water level variations, with maximum salinity occurring at high water. The observed zero lag between salinity peaks, maximum ebb and flood velocities ($u$-component), and water level variations indicate that the tidal propagation is a standing wave.

The vertical velocity profiles showed little shear, except near the bottom as a result of friction. The intensity of the $u$-component of velocity varied semi-diurnally. The up-estuary salt transport was dominated by tidal diffusion. The variation of the Richardson number over time at both sub-estuaries was always below the critical number ($R_i<2$), indicating unstable water column conditions during most of the tidal cycle.

Measuring small-scale physical processes and how they affect the spatial patterns of optical constituents is fundamental for the understanding of the functioning of a coastal ecosystem. Due to the

Fig. 12. Ternary diagrams showing the relative contributions of NAP, phytoplankton and CDOM (based on Prieur and Sathyendranath, 1981) at 443 nm, 555 nm and 675 nm for the Medeiros (black squares) and Itaqui (grey circles) inlets.
influence of freshwater discharge and re-suspension of bottom sediments induced by physical processes, the concentrations of the optical components in the water column generally do not have any simple relationship with bio-optical variables. In some estuaries with higher levels of turbidity the seasonal pattern of absorption and scattering coefficients is driven primarily by changes in the particulate matter of biogenic and mineral origin (GALLEGOS et al., 2005). Distinct scenarios can occur in coastal waters with strong tidal currents, shallow water and important freshwater inputs, such as the English Channel, where meteorology is the main factor driving the bio-optical properties of the water column. The winter period is characterized by the strong dominance of CDOM absorption over the particulate matter, while during the spring-summer period the phytoplankton and CDOM provide a similar contribution (VANTREPOTTE et al., 2007).

The results evidenced the relative contributions of CDOM and the particulate matter, represented by the non-algal particles and the phytoplankton, to light attenuation in the water column. It is clear that the reflectance signal, as observed in other coastal areas (BABIN et al., 2003; STUART et al., 2004; GALLEGOS, 2005; XI et al., 2012; LE et al., 2013), is affected by the complex behaviour of these optically active substances in the study area. Mixed dominances, especially for CDOM and NAP at 443 nm, limit the use of the two-type water type classification model (Cases 1 and 2) commonly used for estuarine regions, such as these study sites. Combining in situ radiometric, bio-optical and biogeochemical measurements can be useful for defining a more comprehensive classification for each specific bio-optical environment, as proposed by VANTREPOTTE et al. (2012).

The terrestrial source of CDOM is likely to be the vast mangrove belts bordering the entire PEC, which strongly suggests that it may be possible to use the optical estimation of CDOM (from the water color) as a proxy for the salinity (XIE et al., 2012, and references therein). Relationships between salinity and $a_{b}g\lambda(\lambda)$, especially in the 400-600 nm band, suggest that the suspension of detritus results from both continental inputs (run-off) and tidal fluxes in shallow waters. The phytoplanktonic community, which relies on mesotrophy (chlorophyll concentrations between 5 and 20 mg·m$^{-3}$, according with the classification of BRICKER et al., 2003), seems to be better developed in higher salinity conditions (>20), which are related to the intrusion of saline waters into the sub-estuaries. The signature of pigments that are believed to be responsible for variations along the phytoplankton absorption spectra, i.e., the shoulders registered at the 625-630 nm band, can be associated with a chlorophyll-c signature (BIDIGARE et al., 1989), indicating the predominance of large diatoms, which are associated with the low $a_{b}g\lambda$ observed in the area (BRICAUD et al., 2004). This result indicates a strong influence of the packaging effect, which may be better explained for the area of the PEC by combining spectral analysis of the phytoplankton absorption with the determination of community composition and density, through the use of microscopy and pigment analysis techniques such as High-performance Liquid Chromatography (HPLC). Because the shoulders at 630 nm black also be associated with phycocyanin, which is an indicator of cyanobacteria (MISHRA et al., 2009), the use of specific pigment band ratios can determine the species composition through the water color and is therefore an important tool for monitoring harmful algae blooms.

Both sub-estuaries presented quite similar behaviour in their hydrographic and hydrodynamic characteristics, with a greater salt transport in the Itaqui due to its larger drainage basin. Despite the few measurements of absorption available, our findings show that the typical typology of complex Case 2 water can be applied to these sub-estuaries. The narrower Medeiros sub-estuary is characterised by higher CDOM influence at the maximum blue absorption, whereas stronger dilution and hydrodynamics with saline waters seem to lead to detritus dominance in the light absorption in the Itaqui sub-estuary. In both systems, saline intrusion is responsible for increasing the phytoplanktonic contribution to the absorption coefficients, especially in the red domain.

Combining the information provided by the in situ observations with bio-optical measurements raises the possibility of determining the spatial patterns of the water circulation in these sub-estuaries using water color images. For instance, despite the high levels of turbulence predominant in these two sub-estuaries, the circulation of water can be described in terms of salinity variations. If any of the optically active substances show conservative behaviour and a strong correlation to salinity, the forecast of tidal-induced patterns of water exchange can be estimated from water color images. Overall, the description of bio-optical properties (using color images) complements the baseline description of these sub-estuaries, and increases the potential for future applications, such as the development/improvement of regional hydrodynamic models.

Overall, the present study brings together an important, albeit limited, baseline description of the hydrodynamics and bio-optical characteristics of two subtropical sub-estuaries, contributing to a better understanding of the influence of forcing mechanisms to the main circulation patterns in such a complex coastal ecosystem that sustains key economic activities in the region.
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